Modelling chemistry in aircraft plumes 1: comparison with observations and evaluation of a layered approach

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Abstract

An expanding plume model with chemistry has been used to study the chemical conversion of nitrogen oxides to reservoir species in aircraft plumes. The model represents the plume by several circular or cylindrical layers in order to give a more detailed description of the chemical evolution in the plume. Model simulations of plumes from two B747 aircraft monitored during the POLINAT measurement campaigns in winter 1994 and summer 1995 were performed. The NOx emission indices were 12.5 and 26.1 g kg⁻¹, respectively. The chemical evolution was followed during the phase dominated by the aircraft-induced dispersion (vortex regime), and in the atmospheric-induced dispersion regime up to 15 h. The results were compared to observations of NO, HNO₃, HNO₂ and CO₂ taken within the trailing vortices. The modelled values were in broad agreement with the observations. With a NOx emission index of 26.1 g kg⁻¹, 60% remained as NOx after 15 h at 50°N under summer conditions for emissions at 07 UT. This amount decreased to 40% when a one-layered plume was considered, suggesting that the distribution of NOx leads to different oxidising potentials which again influenced the plume chemistry. © 2000 Elsevier Science Ltd. All rights reserved.

Keywords: Aircraft emissions; Vortex regime; Dispersion regime; Plume chemistry; Tropospheric chemistry

1. Introduction

Over the last couple of years, substantial work has been done in order to understand the chemical and dynamical processes in aircraft plumes (e.g. Danilin et al., 1994; Dürbeck and Gerz, 1995, 1996; Hanisco et al., 1997; Karol and Ozolin, 1994; Karol et al., 1997; Kärcher, 1995; Kärcher et al., 1996; Meijer et al., 1996; Moulik and Milford, 1999; Möllerhoff et al., 1996; Petry et al., 1998; Schumann et al., 1995, 1998). The emissions of nitrogen oxides (NOₓ = NO + NO₂), hydrocarbons (HC) and sulphur dioxide (SO₂) are gradually converted in the plume, and secondary components, such as e.g. ozone, nitric acid (HNO₃) and sulphate aerosols, are formed. In order to quantify the chemical and climatic impact of aircraft emissions on a global scale, it is of importance to understand the dispersion and chemical transformations that take place in the aircraft plume before it is homogeneously mixed with the surrounding air.

This article is based on work done within the Pollution from Aircraft Emissions in the North Atlantic Flight Corridor (POLINAT)-projects, and presents findings concerning the chemical conversion of NOₓ to reservoir species in aircraft exhaust plumes. Plumes from two different B747s are simulated from a few seconds (s) up to several hours (h) after emission. The model used in our studies describes both the chemical conversions and the dispersion of the exhaust plumes. Similar studies for both subsonic and supersonic aircraft emissions have been performed (Brasseur et al., 1998 and references therein; IPCC, 1999 and references therein). Our approach is to study the chemical evolution in several cylindrical layers, since the plume is not well mixed from
a cross-sectional point of view. The chemical conversions therefore take place at different rates depending on the distance from the core of the plume.

The aims of the present study were twofold. First, the model results were compared to observations of young plumes of some minutes in order to evaluate the dilution of emitted NO\textsubscript{x} and conversion to HNO\textsubscript{3} and nitrous acid (HNO\textsubscript{2}). In principle, we are interested in the evolution of the chemistry in plumes of age up to several hours. The chemistry and dispersion in the jet and vortex regimes do, however, govern the concentration and distribution of the emitted species at the start of the dispersion regime. This affects the chemical evolution in the plume on longer time scales. We modelled two cases of emissions from a B747 with different NO\textsubscript{x} and hydroxyl radical (OH) emission indices (EI). In situ measurements of NO\textsubscript{x}, HNO\textsubscript{3}, HNO\textsubscript{2} and CO\textsubscript{2} mixing ratios in near-field exhaust plumes were available from the POLINAT measurement campaigns on 13 November 1994 and on 30 June 1995. For the summer case, the modelled NO\textsubscript{x} in the plume versus time was also compared to an evolution based on observations suggested by Schumann et al. (1998). The second aim of this study was to investigate the chemical transformation of NO\textsubscript{x} in the different layers in the plume and to quantify the effect of representing the plume by several circular or cylindrical layers versus a homogeneous approach. This was also discussed by Petry et al. (1998) and Karol et al. (1997) and their findings will be compared to our results.

In a companion paper, the impact of varying atmospheric conditions on the chemical evolution in aircraft plumes was studied, and suggestions given for how plume-scale processes can be taken into account in meso-scale model studies (Kraabøl and Stordal, 2000).

2. An expanding plume model with chemistry

The plume behind an aircraft can be described in three regimes: the jet regime, the vortex regime and the dispersion regime. The present model describes the latter two. Calculations start at the end of the jet regime, i.e. after between 1 and 10 s (Hoshizaki et al., 1975). We assume that both at the end of the jet as well as during the vortex regime, the aircraft wakes can be replaced by a single plume with an effective circular cross-section, A. The standard deviation (\(\sigma\)) in the concentration of a passive tracer at the start of the vortex regime was chosen to be 6 m. This leads to mixing ratios in accordance with a dilution rate based on observations (Schumann et al., 1998). The breakdown of the vortex structure was set at 124 s with \(\sigma_{\text{vortex}}(t) = 20 \text{ m}\) determined from near-field plume measurements (Kraabøl et al., 1999). From the subsequent atmospheric-induced dispersion regime the vertical and horizontal standard deviations are set at 50 and 120 m, respectively (Schumann et al., 1995). Atmospheric dispersion is then described using a Gaussian approximation (Konopka, 1995; Schumann et al., 1995):

\[
\begin{align*}
\sigma_{h,0}^2 &= \frac{3}{2} s^2 D_h \Delta t^3 + 2(s_d D_h + s^2 \sigma_h^2) \Delta t^2 + 2(s_d \sigma_h + D_h) \Delta t + \sigma_h^2, \\
\sigma_{v,0}^2 &= s_d D_v \Delta t^2 + (s_d \sigma_h^2 + 2 D_h) \Delta t + \sigma_h^2, \\
\sigma_{z,0}^2 &= 2 D_e \Delta t + \sigma_h^2, \\
\Delta t &= t_{n+1} - t_n, \quad n = 1, \ldots, n - 1,
\end{align*}
\]

where \(\sigma_h, \sigma_v, \sigma_z\) and \(\sigma_h^*\) are the horizontal, vertical and skewed standard deviations, respectively, \(t\) is the time and \(n\) is a time period where the dispersion parameters are held constant. We assume that \(\sigma_{h,0} = 0\) at the start of the dispersion regime. \(D_h, D_v, D_e\) and \(D_e\) are the horizontal, vertical, and skewed diffusion coefficients, respectively, determined according to Dürbeck and Gerz (1996). For the given stratification with a Brunt-Väisälä frequency \(N = 0.014 \text{ s}^{-1}\), \(D_h = 16.7 \text{ m}^2 \text{s}^{-1}\), \(D_v = 1.1 \text{ m}^2 \text{s}^{-1}\) for \(3 < t < 13 \text{ min}\) and \(D_v = 0.15 \text{ m}^2 \text{s}^{-1}\) for \(t > 13 \text{ min}\). Dürbeck and Gerz (1996) found that \(D_h\) was rather small compared to a theoretical limit \((D_h = (D_e D_h)^{1/2})\). In our work, the relation \(D_h = 0.5(D_e D_h)^{1/2}\) is assumed. The mean vertical wind shear, \(s\), perpendicular to the mean plume is 0.005 \text{ m s}^{-1}.

The Gaussian spreading is, strictly speaking, only valid for a passive tracer. In the model application, the aim is to describe the evolution of chemically active species in the plume. The concentration gradients within the plume can influence this evolution, since the chemistry is non-linear. The plume is therefore divided into eight circular or elliptical layers parallel to the centreline in order to resolve the cross-sectional concentration distribution (Fig. 1, lower panel, only three layers are shown). The chemical production and loss in each layer is estimated. The chemical module includes 66 compounds and 173 chemical reactions describing the ozone chemistry in the free atmosphere (Kraabøl et al., 1999). This includes CH\textsubscript{4}, CO, H\textsubscript{2}O, SO\textsubscript{2} and the following oxygen-(O(3P), O(1D), O\textsubscript{3}), hydrogen-(H, H\textsubscript{2}, OH, HO\textsubscript{2}, H\textsubscript{2}O\textsubscript{2}) and nitrogen compounds (NO\textsubscript{x}) (NO, NO\textsubscript{2}, NO\textsubscript{3}, N\textsubscript{2}O\textsubscript{5}, HNO\textsubscript{2}, HNO\textsubscript{3}, PAN (CH\textsubscript{3}COO\textsubscript{2}N\textsubscript{2}), HNO\textsubscript{2}, CH\textsubscript{3}O\textsubscript{2}NO\textsubscript{2}, CH\textsubscript{3}ONO\textsubscript{2}, C\textsubscript{2}H\textsubscript{4}ONO\textsubscript{2}, C\textsubscript{2}H\textsubscript{5}ONO\textsubscript{2}) as well as the hydrocarbons C\textsubscript{3}H\textsubscript{6}, nC\textsubscript{4}H\textsubscript{10}, C\textsubscript{2}H\textsubscript{4}, C\textsubscript{3}H\textsubscript{6}, o-xylene and C\textsubscript{6}H\textsubscript{6}. Several secondary compounds from the oxidation of hydrocarbons are described, which includes the peroxy radicals CH\textsubscript{3}O\textsubscript{2}, C\textsubscript{2}H\textsubscript{5}O\textsubscript{2}, C\textsubscript{4}H\textsubscript{9}O\textsubscript{2}, CH\textsubscript{3}COO\textsubscript{2}. The numerical procedure used is based on the quasi-steady-state approximation (QSSA) (Hesstvedt et al., 1978). Diffusion within the plume is taken into account by estimating the net mass flux of a chemical species through the boundaries of the plume layers. This flux depends on the concentration gradient between the layers, and is further described in Kraabøl et al. (1999).
the plume expands, the boundaries of the layers are redefined and the grid resolution is adjusted, thereby entraining surrounding air and keeping the boundaries in accordance with Eq. (2). The concentrations of the chemical species are redistributed by volume weighting:

\[
c_{k,i}(t + dt) = \frac{c_{k,i}(t)\tilde{V}(t) - \tilde{V}(t + dt) - c_{k,i+1}(t)\tilde{V}(t + dt) - V(t)}{\tilde{V}(t + dt) - \tilde{V}(t)}
\]

where \(c_{k,i}\) is the concentration in layer \(i\) for species \(k\) and \(\tilde{V}\) is the volume at \(t = t + dt\) for layer \(i\). Note \(c_0\) is set equal to the background concentration.

3. Experimental data and initial concentrations

The model described above is applied to data from two flights of the POLINAT measurement campaigns with the DLR’s Falcon research aircraft on 13 November 1994 and 30 June 1995. These two cases will hereafter be referred to as the winter and summer case. The experimental data describing the state of the atmosphere during the chasing of the two B747s, the position of the Falcon, engine type and emission indices are given in Table 1. The emission rate from the B747 in g s\(^{-1}\) is calculated based on the fuel consumption per kilometre, the air speed and the EIs for NO\(_x\), SO\(_2\), water vapour (H\(_2\)O), HC and carbon monoxide (CO) given in Table 1. The distribution of the hydrocarbon emissions is based on suggestions from Pleijel et al. (1993).

The concentrations of oxidants such as OH, HO\(_2\) and O decrease in the turbine and exhaust nozzle as well as during the jet regime due to chemical processes (e.g. IPCC, 1999; Tremmel et al., 1998; Böckle et al., 1999). Kärcher et al. (1996) have studied the chemical evolution in a B747 exhaust plume during the jet regime. They give a detailed description of the plume chemistry, and sug-
Table 1
The state of the atmosphere during the chasing of the plumes from the two B747s and the emissions from the considered B747s. Here, $u_x$, $v_y$, and $z_z$ are related to the west–east ($u$) and south–north ($v$) directions, respectively. True air speed is the speed relative to the wind speed. The given emission indices are stoichiometric or modelled values. The value of $EI(NO_x) = 12.5 \frac{g}{kg}$ was based on near-field plume measurements (Schulte et al., 1997). The data for the chasing on 30 June are taken from Tremmel et al. (1998).

<table>
<thead>
<tr>
<th>Parameter</th>
<th>Units</th>
<th>13 November</th>
<th>30 June</th>
</tr>
</thead>
<tbody>
<tr>
<td>Latitude</td>
<td></td>
<td>51°20'N</td>
<td>53°N</td>
</tr>
<tr>
<td>Longitude</td>
<td></td>
<td>13°1'W</td>
<td>14°30'W</td>
</tr>
<tr>
<td>Assumed emission time</td>
<td>UT</td>
<td>$\approx 14.59$</td>
<td>$\approx 13.55$</td>
</tr>
<tr>
<td>Cruising height $H$</td>
<td>km</td>
<td>11.25</td>
<td>9.2</td>
</tr>
<tr>
<td>Temperature $T_a$</td>
<td>K</td>
<td>214 ± 1</td>
<td>231</td>
</tr>
<tr>
<td>NO (observed)</td>
<td>ppb</td>
<td>0.130</td>
<td>0.0050</td>
</tr>
<tr>
<td>NO$_2$ (observed)</td>
<td>ppb</td>
<td>0.090</td>
<td>0.0020</td>
</tr>
<tr>
<td>HNO$_3$ (observed)</td>
<td>ppb</td>
<td>0.570</td>
<td>0.93</td>
</tr>
<tr>
<td>O$_3$ (observed)</td>
<td></td>
<td>38</td>
<td>118</td>
</tr>
<tr>
<td>Engine type</td>
<td></td>
<td>CF6-80C2B1F</td>
<td>JT9D-7A</td>
</tr>
<tr>
<td>True air speed $V$</td>
<td>m s$^{-1}$, Ma</td>
<td>247, 0.84</td>
<td>247, 0.81</td>
</tr>
<tr>
<td>EI(H$_2$O)</td>
<td>g kg$^{-1}$</td>
<td>1242.0</td>
<td></td>
</tr>
<tr>
<td>EI(CO$_2$)</td>
<td>g kg$^{-1}$</td>
<td>3153.0</td>
<td></td>
</tr>
<tr>
<td>EI(SO$_2$)</td>
<td>g kg$^{-1}$</td>
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<td></td>
</tr>
<tr>
<td>EI(CO)</td>
<td>g kg$^{-1}$</td>
<td>1.5</td>
<td></td>
</tr>
<tr>
<td>EI(NO$_2$)</td>
<td>g kg$^{-1}$</td>
<td>12.5</td>
<td>26.1</td>
</tr>
<tr>
<td>EI(HC)</td>
<td>g kg$^{-1}$</td>
<td>0.2</td>
<td></td>
</tr>
<tr>
<td>Fuel consumption</td>
<td>kg km$^{-1}$</td>
<td>12.0</td>
<td>13.48*</td>
</tr>
</tbody>
</table>

*Calculated based on a true air speed of 247 m s$^{-1}$ and fuel flow of 12 t h$^{-1}$ taken from Tremmel et al. (1998).

The state of the atmosphere during the chasing of the plumes from the two B747s and the emissions from the considered B747s. Here, $u_x$, $v_y$, and $z_z$ are related to the west–east ($u$) and south–north ($v$) directions, respectively. True air speed is the speed relative to the wind speed. The given emission indices are stoichiometric or modelled values. The value of $EI(NO_x) = 12.5 \frac{g}{kg}$ was based on near-field plume measurements (Schulte et al., 1997). The data for the chasing on 30 June are taken from Tremmel et al. (1998).

In this study, the formation and evolution of particles is described. Results from a model describing the aerosol nucleation driven by condensation (Ford et al., 1996) are used as input. The time evolution of particles was calculated with a relative humidity of 43% with respect to liquid water at 214 K. After about 0.1 s the soot particles in the plume are activated and freeze. The supersaturation in the plume reaches ambient levels after about 20 s. Aerosols melt after about 10 s, and the soot particles emitted stabilise at a radius of 0.09 cm with highly acidic compositions. At the start of the calculations, $t = 4$ s, the total surface area was estimated as $5.54 \times 10^4 \mu m^2 cm^{-3}$. After about 1 h it had decreased to $3.5 \times 10^{-5} \mu m^2 cm^{-3}$ and reached the ambient level of soot particles, which is assumed to be from $10^{-5}$ to $10^{-2} \mu m^2 cm^{-3}$ (IPCC, 1999). Reported surface areas in young contrails of age 0.1–0.5 s are from $10^4$ to $10^5 \mu m^2 cm^{-3}$ (IPCC, 1999 and references therein). Kärcher (1997) estimated a total surface area of $6–7 \times 10^4 \mu m^2 cm^{-3}$ after 4 s in the plume.

The same evolution of particles is used in all our calculations, although we are aware that other plume mixing rates and meteorological conditions could cause different types of particle evolution in the plume. After the particles melt, we assume that the soot particles are coated...
with sulphuric acid solution (\(H_2SO_4/H_2O\)) (Kärcher, 1997). The following heterogeneous reaction was assumed to take place on the surface of the ice or coated soot particles:

\[
N_2O_3 + H_2O \rightarrow 2HNO_3. \tag{6}
\]

For coated soot particles the reactive uptake coefficient for sulphate particles was used, and set to 0.1 in accordance with JPL (1997).

4. Results

4.1. Comparison with observations

The model was applied to the two cases described in Section 3, and compared to near-field measurements. The division of the plume into several layers causes an increasing concentration gradient of the emitted species towards the centre of the plume. As the concentration of the emitted species in the various layers of the plume can differ by an order of magnitude, accurate comparison with observed data is difficult, especially since we do not know exactly where the Falcon penetrated the plumes. Our aim was therefore not to reproduce the near-field observed values, but evaluate the magnitude of difference between the modelled and observed values.

Chemical instrumentation on board the Falcon for the aircraft exhaust plume measurements included a chemiluminescence (CL) detector for NO, a differential non-dispersive infrared (NDIR) analyser for \(CO_2\), and a chemical ionization mass spectrometer (CIMS) for \(HNO_2\) and \(HNO_3\). Details on the three instruments and measuring techniques have been described by Schlager et al. (1997), Schulte et al. (1997), and Arnold et al. (1992), respectively. Briefly, the NO detector uses \(NO/O_2\)–CL technique including a zero volume upstream of the reaction vessel. Sample air is taken from outside the aircraft via a Teflon inlet line. The pressure in the sampling line is kept constant at 120 hPa by pumping excess air through a pressure-controlled Teflon-valve. The CL reaction chamber is operated at a constant sample gas flow of \(31 \text{ min}^{-1} \) (STP) and pressure and 35 hPa. For calibration the instrument is supplied with sample air of known NO mixing ratios using BOC standard \(NO/N_2\) mixtures diluted with zero air. The NDIR instrument includes two absorption cuvettes, an infrared source, and a pneumatic infrared detector. Measuring air is sampled via a Teflon line from outside the aircraft and compressed to 1.1 atm by a diaphragm pump before being passed through one of the cuvettes. The second cuvette is flushed with a reference mixture of \(380 \text{ ppmv} \ CO_2\) in dry air. For calibration the instrument is supplied with pre-prepared \(CO_2/air\) standards. In the CIMS instrument, \(HNO_2\) and \(HNO_3\) are ionised by specific ion–molecule reactions in a flow tube attached to a quadrupole mass spectrometer. The mass spectrometer is pumped by a liquid-neon-cooled cryopump. The reactant ions are produced by a high-frequency glow discharge and injected into the flow tube. The \(HNO_2\) and \(HNO_3\) abundances are calculated from the measured ratios of product and reactant ions, the reaction rate coefficient, and the ion residence time in the flow tube. The estimated accuracies for the measurements of \(CO_2\), NO, \(HNO_2\), and \(HNO_3\), enhancements in the aircraft exhaust plumes are 8, 6, 50, and 50\%, respectively.

The aircraft plume measurements were performed off the west coast of Ireland in the Shannon Radar Control Zone for the transatlantic air traffic. The chasing procedures of specific source aircraft are described by Schulte et al. (1997). During the chasing of a B747 on 13 November 1994, maximum NO concentrations were measured in the two vortices, which descended about 150 m from the cruising altitude of the source aircraft due to wake dynamics. The remaining NO data were mostly obtained from the so-called secondary plume, a diffuse curtain of emissions which were detrained or originally not trapped by the vortices. For comparison with the plume model simulations, only the observations in the vortices are used.

Fig. 3 shows the modelled evolution of NO in the plume of the B747 during the first 4 min for the winter case. When the vortex structure breaks down, the mixing ratio dropped in all the layers of the plume due to a rapid plume expansion causing a high entrainment of surrounding air. During the first 3 min, NO in the innermost layer decreased from 427 to \(\sim 4 \text{ ppbv}\). The Falcon only penetrated the trailing vortices once, when the plume was 83 s old. The measured NO was \(59 \pm 5 \text{ ppbv}\) and the modelled values in the five inner layers of the plume ranged from 112 to 42 ppbv. The modelled values in
the middle part of the plume were in agreement with the observed NO.

For the summer case, the vortices were penetrated three times. Fig. 4, left panel, shows the NO value in the various layers of the plume together with observations during the first 4 min. The modelled values in the four inner layers ranged from 264 to 143 and 203 to 110 ppb after 59 and 75 s of calculations, respectively. This was in reasonable agreement with the observed maximum of 184 ± 15 and 158 ± 12 ppb at the same times.

For both comparisons, the modelled values in the innermost part of the plumes are high compared to the observations. These observations are taken within the trailing vortices, but normally such observations were avoided due to high turbulence (Schulte et al., 1997). Since we do not know if the penetrations reached the core of the plume, we allow a somewhat higher concentration in the inner layers. Schulte et al. (1997) report nighttime measurements of NO concentration up to 318 ppb, between 61 and 104 s after emissions associated with an EI(NO$_x$) of 17 g kg$^{-1}$ and a fuel flow ratio of 10.4 t h$^{-1}$.

This might indicate that higher plume values than 184 ppb, which were measured, could have occurred given the aircraft EI(NO$_x$) of 26.1 g kg$^{-1}$ and fuel flow ratio of 13.6 t h$^{-1}$.

For the summer case, we have compared the modelled NO$_x$ mixing ratio versus time with a time evolution suggested by Schumann et al. (1998),

$$\delta r_{NO_x} = \frac{EI(NO_x)M_{air}}{M_{NO_x} \cdot N}, \quad N = 7000 \left( \frac{t}{t_0} \right)^{0.8},$$

where $\delta r_{NO_x}$ is the mixing ratio of NO$_x$, $M_{air}$ and $M_{NO_x}$ the molecular weight of air and NO$_2$, respectively. $N$ is the bulk dilution rate. Fig. 4, right panel, shows the change in volume mixing ratio of NO$_x$ in the inner and outer parts of the plume at approximately 3 h after emission. In the vortex regime, the modelled concentration in the inner part of the plume was about five times higher than the concentration calculated by Eq. (7) (solid line). As pointed out by Schumann et al. (1998), $N$ is the bulk dilution ratio based on measurements in more than 70 plumes. The fit is best for young plumes of age from 1 to 50 s, and may deviate by a factor of 3–5 from Eq. (7).

The evolution suggested by Schumann et al. (1998) was, however, in fair agreement with the time evolution of the modelled mean NO$_x$ in the plume (diamonds). The highest deviation was found after the breakdown regime (3 min after emission). As discussed in Schumann et al. (1995), the vortex regime can last from 3 to 5 min. If we delay the modelled breakdown by 2 min, the mean NO$_x$ value in the plume was 16 ppb after 3 min. This is within the uncertainty limit given for Eq. (7). We found, however, that allowing a longer vortex period did not have any significant impact on the amount of NO$_x$ converted to reservoir species on longer time scales.

Fig. 5 shows the observed values of HNO$_3$ and HNO$_2$ together with modelled data for the summer case. As seen for NO, the modelled HNO$_2$ in the middle part of the plume was in agreement with the observed value of 2.6 ± 1.6 and 2.3 ± 1.1 ppb after 59 and 75 s, respectively.

The measured CO$_2$ concentrations above the background in the vortices were 25.4 ± 1.5, 21 ± 1.3 and 4.8 ± 0.3 ppm at 59, 75 and 86 s, respectively (not shown). The background concentration was 362 ppm (Tremmel et al., 1998). The modelled values above the background in the four inner layers of the plume at the same times ranged from 40.4 to 22.2, 30.5 to 16.7 and 25 to 16.7 ppm.

These comparisons indicate that the model is able to estimate reasonable values of NO, HNO$_2$, HNO$_3$ and...
CO₂ with respect to the concentrations measured in the vortices of young aircraft plumes.

4.2. Chemical conversion to reservoir species in the different layers of the plume

The aircraft plume in the summer case, was followed for 15 h. We assume that the plume was emitted at 07 UT at 50°N and 0°W in July. The temperature was set to 230 K. The NOₓ conversion to reservoir species like HNO₃, methyl-peroxynitrate (CH₃NO₄), PAN and peroxynitric acid (HNO₄) in the different layers of the plume was investigated for the 15 h period. Table 2 lists concentrations of some key species in the plume and in the background air at noon and midnight. Diurnal variations of the concentrations in the background air are taken into account in the model calculations.

To study the role of the chemistry, two model runs were compared: a run with emissions (A) and a run without emissions (B). In run B, we follow an imaginary plume with the same dispersion and mixing with ambient air as in run A. The purpose was to simulate the natural chemical diurnal variations along the plume trajectory. In this case, the concentration in the ‘plume’ is very similar to the concentration in the ambient air, but can differ if the plume is advected over large distances. The emission rate of NOₓ was 81.6 g s⁻¹ and this gives 1.05 × 10²⁴ NOₓ molecules emitted at the engine exit. By studying the decrease in the emitted NOₓ and the corresponding increase in reservoir species, it is possible to quantify the amount of the emitted NOₓ, which is converted to reservoir species versus time. For this purpose, the following equation was used:

\[ f(t)_{k,i} = V(t) [c(t)_{k,i(A)} - c(t)_{k,i(B)}], \]  

where \( f(t)_{k,i} \) gives the difference in numbers of molecules between the runs for a species \( k \) in a layer \( i \) versus time \( t \). Suffixes A and B refer to runs with and without emissions, respectively. Mixing with ambient air is the same for both runs, so that \( f(t)_{NO} \) for the whole plume is the sum of emitted NOₓ, and is constant in time. By studying e.g. the fraction \( \sum f(t)_{HNO₃,i}/f(t)_{NO} \) or \( f(t)_{HNO₄,i}/f(t)_{NO} \), it is possible to quantify the fraction of the emitted NOₓ which is chemically converted to HNO₃ versus time in the whole plume or in a layer. This approach was used in the results shown in Fig. 6, Tables 3 and 4.

It is important to stress that the results shown are not sensitive to the chosen length \( L \) of the plume segment (Fig. 1). For a given emission index, fuel consumption and aircraft speed, a shorter length gives a lower emission rate, but the initial volume of the plume will also decrease.

### Table 2

<table>
<thead>
<tr>
<th>Initial concentration</th>
<th>Plume</th>
<th>Background Noon/midnight</th>
</tr>
</thead>
<tbody>
<tr>
<td>NO (ppb)</td>
<td>730</td>
<td>0.034</td>
</tr>
<tr>
<td>NO₂ (ppb)</td>
<td>98</td>
<td>0.012/0.046</td>
</tr>
<tr>
<td>HNO₃ (ppb)</td>
<td>3.1</td>
<td>0.32/0.33</td>
</tr>
<tr>
<td>OH (ppt)</td>
<td>0.40</td>
<td></td>
</tr>
<tr>
<td>HO₂ (ppt)</td>
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</tr>
<tr>
<td>O₃ (ppb)</td>
<td>85.2</td>
<td>85.4/85.6</td>
</tr>
<tr>
<td>PAN (ppb)</td>
<td>0.25</td>
<td>0.25/0.26</td>
</tr>
<tr>
<td>HCHO (ppb)</td>
<td>1.18</td>
<td>0.056/0.060</td>
</tr>
<tr>
<td>CO (ppbm)</td>
<td>2.56</td>
<td>113/113</td>
</tr>
<tr>
<td>C₂H₄ (ppt)</td>
<td>2.37</td>
<td>1.43/0.82, 19⁵</td>
</tr>
<tr>
<td>C₂H₆ (ppb)</td>
<td>2.10</td>
<td>2.04/2.03</td>
</tr>
<tr>
<td>C₃H₆ (ppb)</td>
<td>0.48</td>
<td>0.002⁵</td>
</tr>
<tr>
<td>NC₄H₁₀ (ppb)</td>
<td>0.8</td>
<td>0.028⁶</td>
</tr>
<tr>
<td>o-xylene (ppb)</td>
<td>0.12</td>
<td>0.002⁶</td>
</tr>
<tr>
<td>C₃H₈ (ppt)</td>
<td>6.3</td>
<td></td>
</tr>
</tbody>
</table>

*The diurnal monthly mean concentrations of short-lived hydrocarbons are given, since the concentrations were reduced during the modelled spin-up time.
Fig. 6. Left panel: The estimated chemical conversion of the emitted NOx (26.1 g kg\(^{-1}\)) to the different species in the odd nitrogen family versus time. Right panel: The percentage conversion to HNO\(_3\) in the different layers of the plume versus time.

### Table 3

The amount of the emitted NO\(_x\) remaining as NO\(_x\) after 15 h in a layered and a homogeneous plume for different background concentrations of NO\(_x\) and ozone

<table>
<thead>
<tr>
<th>NO(_x) (ppt)</th>
<th>Ozone (ppb)</th>
<th>Layered (%)</th>
<th>Homogeneous (%)</th>
<th>Layered (%)</th>
<th>Homogeneous (%)</th>
</tr>
</thead>
<tbody>
<tr>
<td>50</td>
<td>150</td>
<td>65</td>
<td>51</td>
<td>60</td>
<td>33</td>
</tr>
<tr>
<td>150</td>
<td></td>
<td>67</td>
<td>51</td>
<td>66</td>
<td>42</td>
</tr>
</tbody>
</table>

### Table 4

The same as Table 3, but for different times of emissions. The NO\(_x\) and ozone background concentrations were 150 ppt and 150 ppb, respectively.

<table>
<thead>
<tr>
<th>Time (UT)</th>
<th>Layered (%)</th>
<th>Homogeneous (%)</th>
</tr>
</thead>
<tbody>
<tr>
<td>1</td>
<td>57</td>
<td>41</td>
</tr>
<tr>
<td>5</td>
<td>60</td>
<td>39</td>
</tr>
<tr>
<td>9</td>
<td>72</td>
<td>50</td>
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<tr>
<td>13</td>
<td>84</td>
<td>72</td>
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<tr>
<td>17</td>
<td>80</td>
<td>79</td>
</tr>
<tr>
<td>21</td>
<td>57</td>
<td>56</td>
</tr>
</tbody>
</table>

since the aircraft covers a shorter distance. The concentration at the start of the calculations is therefore not influenced by the length of the plume.

Fig. 6, left panel, shows the distribution of the NO\(_x\) emissions with time for the plume seen as a whole. After 15 h, 40% of the emitted NO\(_x\) was converted to reservoir species, mainly HNO\(_3\). The highest conversion was calculated in the two inner layers of the plume during most of the 15 h period (Fig. 6, right panel).

The second most important reservoir was CH\(_3\)NO\(_4\). This compound is formed by reaction between CH\(_3\)O\(_2\) and NO\(_2\), and is thermally unstable. The temperature used in this case was 231 K, and the lifetime of CH\(_3\)NO\(_4\) was about 6 h. Reaction- and photolysis rates are adapted from Atkinson et al. (1997) and JPL (1997), respectively. After 15 h, 2% of the NO\(_x\) was converted to this compound. The conversion to HNO\(_4\) by the reaction between NO\(_2\) and HO\(_2\), also amounted to about 2%. This conversion was faster in the outer parts of the plumes. The reaction

\[
\text{NO} + \text{HO}_2 \rightarrow \text{NO}_2 + \text{OH},
\]

caused a stronger decrease in HO\(_2\) in the inner layers due to higher NO. The conversion to PAN was small in all layers due to a limited availability of the CH\(_3\)COO\(_2\) needed for PAN conversion. Furthermore, the build up of dinitrogenpentoxide (N\(_2\)O\(_5\)) was low in all the layers after 15 h. N\(_2\)O\(_5\) is easily dissociated by UV-radiation and is not a reservoir for NO\(_x\) during daytime. The conversion of NO to HNO\(_2\) in the jet regime was set to 1.3%. After 15 h, the fraction had decreased and the loss of HNO\(_2\) was higher than the production.

Möllhoff et al. (1996) used a plume model to study emissions at 50°N under June conditions. The NO\(_x\) emission index was 21.2 g kg\(^{-1}\). For emissions at 08 local time they
calculated the following distribution after 15 h, 70% as NO\textsubscript{x}, 23% as HNO\textsubscript{3}, 5% as HNO\textsubscript{4} and 2% as N\textsubscript{2}O\textsubscript{5} (based on Fig. 3, right panel in Möllhoff et al., 1996). We used the same emission time, temperature (214 K) and a background concentration of NO\textsubscript{x} of 150 ppt, and estimated 75% as NO\textsubscript{x} and 24% as HNO\textsubscript{3} after 15 h. The total conversion to reservoir species is comparable, but in this work the build-up of HNO\textsubscript{3} was lower. This might be due to different values of HO\textsubscript{2} in the ambient air. As explained above, the conversion to this compound was faster in the outer layers of the plume, and is therefore strongly influenced by the concentrations in the surrounding air.

Meijer et al. (1996) used a layered plume model and estimated the following distribution after 15 h for emissions at noon at 50°N under July conditions; 75% as NO\textsubscript{x}, 15% as HNO\textsubscript{2} and 10% as N\textsubscript{2}O\textsubscript{5}. The NO\textsubscript{x} background was 50 ppt and the NO\textsubscript{x} emission index was 16 g kg\textsuperscript{-1}. For the same conditions, but with EI(NO\textsubscript{x}) = 12.5 g kg\textsuperscript{-1}, our modelled distribution was 77% as NO\textsubscript{x}, 20% as HNO\textsubscript{3}, 0.5% as N\textsubscript{2}O\textsubscript{5} and 3.5% as HNO\textsubscript{4}. As in the first comparison, the modelled conversion of NO\textsubscript{x} was in good agreement, but with less build-up of N\textsubscript{2}O\textsubscript{5}. The reason for the discrepancy could be the different turbulent conditions used in the two studies. Meijer et al. (1996) calculated the dispersion of the plume using a Gaussian approximation, but did not include vertical shear. In our studies, vertical shear implies a higher plume dispersion rate and, hence, a faster dilution of the emitted NO\textsubscript{x}. For emissions at noon, the NO\textsubscript{x} concentration will be lower in the plume during night in our study and this might explain the reduced conversion to N\textsubscript{2}O\textsubscript{5}.

The chemical conversion to HNO\textsubscript{3} in the different layers is dependent on the concentrations of OH and NO\textsubscript{x}. Fig. 7, left panel, shows the evolution of OH in the five innermost layers of the plume and in the background during the 15 h calculation. OH at the engine exit was assumed to be converted to HNO\textsubscript{2}, HNO\textsubscript{3} and H\textsubscript{2}O\textsubscript{2} during the jet regime. Although a homogeneous distribution of the emissions was assumed at the start of the vortex regime, the evolution of OH was different in the various layers. The maximum rise in OH above the background occurred in layer 2 about 6 h after emissions. OH was 60% higher than the value in the background air. In layer 1, the OH value was below background for the whole period. The concentration started to increase about 1 h after emissions due to both photochemical production and mixing with ambient air. After 12 h the sun set, and this caused a strong decline in OH concentration in both the plume and the background air.

The evolution of OH in the plume was closely linked to the evolution of ozone in the different layers (Fig. 7, right panel). At the start of the calculations, the ozone level dropped in all the layers due to reaction with NO. Fig. 8 shows that the highest NO\textsubscript{x} concentration was estimated in the innermost layer of the plume, and this caused the most pronounced ozone drop. Due to both mixing with ambient air and chemical production, the levels thereafter increased differentially in the layers. In the innermost layer, ozone reached a maximum about 12 h after emissions. There was probably a potential for higher ozone, but with sun set, ozone concentrations approached background. The most efficient ozone production was calculated in layer 2, where the highest OH increase also was calculated (Fig. 7).

The heterogeneous reaction (6) had little influence on the conversion of NO\textsubscript{x} to reservoir species, and thereby the ozone chemistry in the plume as reported elsewhere (e.g. Moulid and Milford, 1999; Danilin et al., 1994; Karol and Ozolin, 1994).

Petry et al. (1998) calculated effective emission indices of reservoir species in aircraft plumes using a Gaussian approach. In contrast to our findings, they found that dividing the plume into different layers did not have a significant effect on the conversion of the emitted NO\textsubscript{x}.

![Fig. 7](image-url)
or on the ozone concentration in the plume. Karol et al. (1997) used both a box model and a Gaussian plume model to study the chemistry in exhaust plumes and their findings support Petry et al. (1998). In both works the emissions are assumed to take place at 50°N at summer conditions as in the work reported here. When using a homogeneous approach, we find that 42% of the NOx emissions remained as NOx after 15 h, while this amount increased to 60% in the layered plume. Fig. 8 shows the evolution of NOy in the homogeneous plume (solid lines with squares). The mixing ratio of NO was lower compared to the inner part of the layered plume, causing a slower ozone titration with NO compared to these layers. The ozone production was more efficient in the homogeneous plume, and the maximum ozone increase above background was estimated about 2 h after emissions (Fig. 7, right panel, solid lines with squares). This again favoured OH production in the plume. Fig. 7 shows that the OH level in the homogeneous plume was higher than the background for a longer time period than what was the case in the layered plume. This caused a higher conversion to HNO3 compared to the layered plume. In addition, the conversion to HNO2 was also higher in the former case. A smaller amount of HO2 was lost in reaction with NO compared to the layered plume.

We have also compared a layered versus a homogeneous approach for different background concentrations of NOx and ozone (Table 3). In both approaches, the chemical conversion of NOx to reservoir species increased with a higher ozone concentration in the ambient air. In addition, the difference in the chemical conversion of NOx between the two approaches was enhanced. With a high ozone concentration, the conversion of NOx to HNO3, PAN and CH3NO3 increased in the plume. At the start of the calculations, NO reacts with ozone and converts more of the emitted NO to NO2, when background ozone is higher. Since HO2, CH3COO2 and CH2O2 can all react with NO, this conversion will reduce the decrease in these compounds which are needed for HNO4, PAN and CH3NO4 production, respectively. This effect was more distinct when a homogeneous plume approach was used.

Table 4 lists the results from homogeneous and layered plume calculations for different times of emissions. The difference in the chemical conversion of NOx after 15 h showed a strong diurnal variation. The greatest difference was found for emissions in the morning, and the least for emissions in the afternoon.

It should be noted that the time the plume is followed strongly affects the difference between a layered and a homogeneous approach. Like Karol et al. (1997), we find little difference between the two approaches after 1 h of calculations. In our case, the conditions for ozone production in the plume govern the magnitude of difference between the approaches. The potential for ozone production during daylight is dependent on the ratio between NOx and non-methane hydrocarbons (NMCH) in the plume. A reason for the differences in our work compared to Petry et al. (1998) and Karol et al. (1997), could be a different distribution or amount of the hydrocarbons emitted, or a higher concentration in the ambient air. Moulid and Milford (1999) found that the ozone production in the plume was strongly influenced by the NMCH concentration in the ambient air. How the hydrocarbons are described and their lifetime in the chemistry schemes used in different plume models, will also influence the potential for ozone production.

5. Conclusions

Plumes emitted from two B747 aircraft have been modelled, and the calculated NO, CO2, HNO3 and HNO2 were compared to near-field observations from the POLINAT database. The modelled NOx mixing ratio on longer time scales was compared to a time evolution based on observations in more than 70 aircraft plumes (Schumann et al., 1998). The modelled NO and CO2 values in the innermost layers in the plume were from 90 to 30% and 70 to 60% higher than the observed values in the trailing vortices, respectively. As discussed in Schulte et al. (1997), observations within the trailing vortices were normally avoided due to high turbulence. It is therefore uncertain whether the highest possible NO and CO2 concentrations were measured. The modelled values were therefore in fair agreement with the observations.

In the model application, the main focus has been the conversion of the emitted NOx to reservoir species, such as HNO3, CH3NO4, HNO4, PAN and N2O5. Contrary to Petry et al. (1998) and Karol et al. (1997) we found differences between a layered and a homogeneous approach. In our case, the inhomogeneous distribution of the emitted species caused different oxidising capacity in
the various layers of the plume. At 50°N under summer condition for emissions at 7 UT, 60% of the emitted NOx remained after 15 h. When considering a homogeneous plume, this amount was 43%. In the latter case, the conditions for ozone production in the plume were more favourable, and the conversion of NOx to HNO3 was higher. In addition, a higher conversion of NOx to CH3NO3 was estimated compared to the layered approach. In our study, the difference in the chemical evolution between a layered versus a homogeneous plume, during a 15 h period, strongly depends on the time of emissions and background concentrations of ozone in the ambient air. Nevertheless, we found that the division of the plume into several layers revealed different conditions for the oxidising potential of NOx in the plume. For the following work we therefore use a layered approach when investigating the chemical evolution in aircraft plumes.

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References


